



Test methods for CO₂ mineralization in cement-based materials: A review by RILEM TC 309-MCP

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Abstract CO₂ can be absorbed and mineralized as solid carbonates in a range of construction materials. Accurate measurement of CO₂ content and uptake in cement-based and other construction products is important for the development as well as the certification of novel mineral carbonation products, processes, and practices. However, measurement is complex as samples can range from fine powders to blocks with varying levels of heterogeneity, different solid carbonate phases can be present, and measurement methods may require correction for interferences. When and where the CO₂ content is measured is also important. Numerous techniques are available for measuring the solid CO₂ content of a sample, from which the CO₂ uptake can be calculated. Here, a critical review of the most relevant methods is presented; the most used methods are thermogravimetric and combustion analysis, however, diffraction-based, spectroscopic, wet-chemical, and other methods can also be used for this purpose. Advantages and disadvantages of the various test methods are discussed. Aspects of sample preparation, result interpretation, measurement limitations and possible interferences, as well as the use of complementary methods are

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highlighted. Important aspects of conversion of measured solid CO₂ contents to calculated CO₂ uptake are highlighted.

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1 Introduction

There is great interest worldwide in reducing construction-related CO₂ emissions. The construction materials community is increasingly interested in the important area of mineral carbonation to treat or produce building materials with strongly reduced CO₂ footprint. In mineral carbonation processing, CO₂ is bound in solid products and forms stable alkaline carbonates, mostly CaCO₃ polymorphs. As the CO₂ is permanently bound and intentionally incorporated into building products, this is considered a form of carbon capture and utilization (CCU), and helps to offset CO₂ emissions within or even beyond the production process. A wide range of mineral carbonation building products are being considered, from finely divided powders to be used as supplementary cementitious materials or fillers for cement and concrete; carbonation cured recycled concrete aggregates; carbonation hardened blocks, bricks or pavers, and more. For each of these products, accurate quantification methods for bound CO₂ content and CO₂ uptake are of utmost importance in both scientific and industrial contexts, for instance to provide a sound scientific foundation for CO₂ reduction accounting as well as for certification, carbon credits and emission trading schemes. A wide variety of analytical methods exist that enable to follow various characteristics of the carbonation process. Thermogravimetric analysis, combustion analysis, Fourier transform infrared spectroscopy, X-ray diffraction (XRD), and wet-chemical analysis methods are all established techniques for studying mineral carbonation processes. Yet, due to inherent limitations associated with a given test method, multiple methods are often combined to obtain reliable results or for verification. For each individual method, factors such as sample pretreatment, sample preparation, and signal interferences or overlapping play an important role, and must be well-understood to obtain accurate results. In addition, there are method-specific instrument settings and set-ups, as well as data treatment and result calculation steps that must be taken into account correctly. To support the further development and introduction of mineral carbonation products, establishment of detailed test procedures for accessible measurement methods is of key importance. As a first step, this paper critically reviews and identifies the most promising measurement methods for CO₂ content

and uptake analysis of construction materials and discusses influencing factors, limitations and interferences for each method. We focus largely on calcite, vaterite, aragonite, and amorphous calcium carbonate. Monohydrocalcite, ikaite, magnesium carbonates, iron carbonates, and other related materials are only briefly mentioned as they are relatively uncommon in the carbonation of construction materials. While we are concerned with CO₂ content, we do not focus on CO₂ profile or depth determination. Thus, techniques such as X-ray tomography (both laboratory and synchrotron), neutron imaging, pair distribution function, electron microscopy, and several others are not covered. Finally, this review covers the most commonly used test methods for CO₂ content/uptake determination; less common and emerging methods are not discussed.

2 Thermogravimetric analysis

Thermogravimetric analysis (TGA) is one of the most commonly used methods for CO₂ content measurement. In TGA, a powdered specimen is heated to high temperatures, typically >900 °C, while the specimen mass is monitored precisely. The powder mass is usually around 10–100 mg, but some equipment suppliers allow to go up to 5 g. The solid is ground down to a specified particle size, typically below 45–63 μm, placed in an inert crucible, and loaded into the instrument. Heating rates are commonly 5–20 °C/min, meaning the test takes 45–180 min, with a further 30–60 min duration for the furnace to cool down. As the temperature increases, various phases decompose, losing H₂O or CO₂, and the mass loss and stoichiometry are used to quantify the amount of the phase present. The mass loss gives the amount of CO₂, while the mass loss converted based on stoichiometry gives amounts of CaCO₃ or other carbonate phases. The technique relies on the fact that specific phases have a well-defined, known composition and that specific decomposition reactions occur at well-known temperature ranges. Table 1 presents reaction, dissociation temperature, and mass changes for the most important carbonate phases.

While Table 1 lists ranges, these ranges are broad and are influenced by the temperature ramp rate, the



Table 1 Reaction, dissociation temperature, and mass changes for selected carbonate phases

| Material | Reaction | Temperature | Mass change (g/g) | References |
|--|--|--|------------------------|------------|
| Crystalline calcium carbonate (calcite, vaterite and aragonite) CaCO ₃ | CaCO ₃ → CaO + CO ₂ | 650 – 850 °C | 0.440 | [1, 2] |
| Amorphous calcium carbonate CaCO ₃ ·nH ₂ O | CaCO ₃ ·nH ₂ O → CaO + CO ₂ + nH ₂ O | < 200 °C (H ₂ O)* 480 – 750 °C (CO ₂) | ~0.150–0.200 ~0.340 | [1] |
| Monohydrocalcite CaCO ₃ ·H ₂ O | CaCO ₃ ·H ₂ O → CaO + CO ₂ + H ₂ O | 170 – 380 °C (H ₂ O) 600 – 750 °C (CO ₂) | 0.180 0.370 | [3] |
| MgCO ₃ | MgCO ₃ → MgO + CO ₂ | 400 – 700 °C | 0.522 | [4] |
| FeCO ₃ | 3FeCO ₃ → Fe ₃ O ₄ + 2CO ₂ + CO | 450 – 550 °C | 0.334 | [5] |
| FeCO ₃ | 2FeCO ₃ + 1/2O ₂ → Fe ₂ O ₃ + 2CO ₂ | 450 – 550 °C | 0.311 | [5] |
| NaHCO ₃ | 2NaHCO ₃ → Na ₂ CO ₃ + CO ₂ + H ₂ O Na ₂ CO ₃ → Na + CO ₂ + 1/2O ₂ | 90 – 150 °C | 0.369 | [6] |
| KHCO ₃ | 2KHCO ₃ → K ₂ CO ₃ + CO ₂ + H ₂ O K ₂ CO ₃ → K + CO ₂ + 1/2O ₂ | 100 – 160 °C | 0.310 | [7] |

*Additional water may be included which dehydrates at <200 °C

particle size distribution, the gas composition, amorphous vs. crystalline phases, and the nature of the polymorph present. As an example, the data reported by Radha et al. [1] suggest a significant decrease in the temperature of decomposition of CaCO₃ by 80–120 °C when the specific surface area is increased from 1 to 10 m²/g. The exact choice of temperatures to determine the CaCO₃ content and various corrections can become quite complex [8, 9]. Step, tangential and other mass loss integration methods have been proposed, as have corrections for buoyancy [8, 9]. While quantification of amounts of CaCO₃ as low as 1% appears to be possible, relative errors are expected to be significant at low contents.

Arguably, in most mineral carbonation products, we are limited to CaCO₃, MgCO₃ and intermediates thereof, which form under economically accessible conditions. However, the presence of multiple CaCO₃ polymorphs and the possible occurrence of amorphous CaCO₃ (which decarbonates at lower temperatures between 480 and 750 °C) may complicate the mass loss measurement. Additionally, amorphous calcium carbonates may contain up to 20 wt. % water, which is lost in a temperature range between 50 and 200 °C [1]. The main magnesium carbonate phases occurring in aqueous or moist carbonation are hydrocalcite (Mg₆Al₂CO₃(OH)₁₆·4H₂O and similar compositions with a variable Mg/Al ratio) or hydromagnesite (Mg₅(CO₃)₄(OH)₂·4H₂O). These present CO₂

mass losses below 600 °C, which are intermixed with water and other losses and are difficult to quantify. Using a coupled TGA and mass spectrometry instrument allows to better separate out contributions of amorphous calcium carbonates to mass loss.

TGA suffers from various limitations that are common to all powder-based methods, such as sampling and interference and that are not related to the measurement itself. The main method-related limitation is related to mass loss overlap with other phases, as described above, and precise quantification can be challenging. Various CaCO₃ polymorphs cannot be easily differentiated using TGA and other techniques must be used. By its nature, the technique is destructive, and the sample cannot be reused. Use of coarse particles in testing may cause underestimation of carbonates in cases where carbonates are encapsulated by other materials that can be decomposed at higher temperatures [10]. The coexistence of silica gel and carbonate introduces the Hedvall effect, resulting in the decomposition of calcium carbonate at lower temperature ranges, complicating measurement [11].

TGA instruments are widely available in laboratories, and the TGA is arguably (currently) one of the two most common methods to determine the mineralized CO₂. However, as described above, due to overlaps in decomposition temperatures and the general complexity of cementitious materials, complementary methods are often needed. X-ray



fluorescence spectroscopy enables assessing the precursor chemical composition and can be used to identify the carbonatable components (CaO, MgO, FeO) and helps to estimate the theoretical maximum CO₂ binding potential. Methods such as XRD are also commonly used to complement TGA; through these methods, presence/amounts of phases such as Ca(OH)₂, CaCO₃, MgCO₃ etc., can be determined, which leads to understanding of interferences and further confidence in the measured values of CaCO₃ and MgCO₃. Additionally, TGA can be coupled with a mass or infrared spectrometer to quantify the amounts of released gasses at a certain temperature range, thus correctly allocating CO₂ related mass loss and allowing to differentiate from other sources of mass loss, e.g. water and sulfate.

Finally, when TGA is unavailable, decarbonation mass losses can be measured using step-wise heating in a lab furnace, like those done in loss of ignition (LOI) test. Ideally, such a test would be done with inert gas flow, and the mass loss over the decarbonation temperature range, e.g., between 550 and 850 °C, can be used to estimate CO₂ or CaCO₃ amounts. Such an approach obviously suffers from interferences such as powder instability before or during weighing in a conventional lab balance, and the lack of a mass loss curve makes it less reliable, but it is more practically and widely applicable. This method has the advantage of allowing the use of a much larger sample mass (a few grams) than that of TGA (a few milligrams). The furnace can also be equipped with a balance for continuous monitoring of the sample mass change [12]. It is then similar to TGA but at large scale. Maruyama et al. [13] describe the development of a large-scale thermogravimetric and gas analyzer that can be used to determine CO₂ in concrete. While sample preparation was challenging, the results obtained were in accordance with theoretical values and the experimental duration was relatively short, overcoming some of the challenges described in this section. The throughput and accuracy of TGA for use in larger scale research projects has been reported to be improved by using multi-sample TGA equipment that can test 18 samples at one time with sample weights of 3–4 g [14].

3 Combustion analysis (also known as carbon sulfur (CS) analyzer)

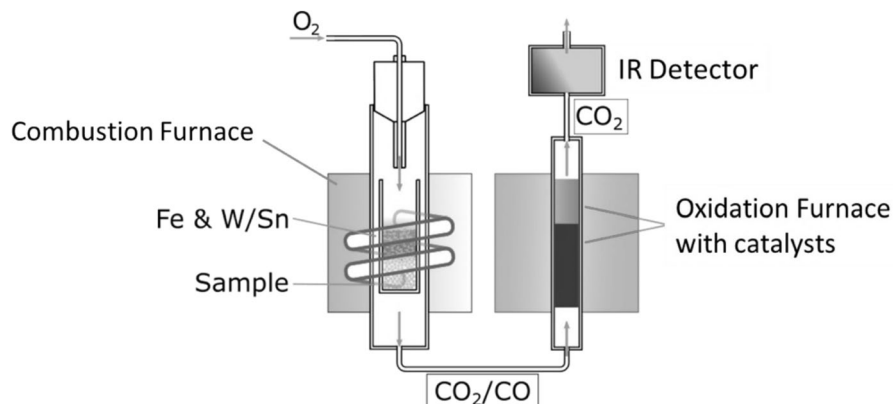
Modern-day combustion analysis is used to quantify carbon and sulfur in a wide variety of materials, including metals, metal alloys, cements, and other materials [15, 16]. The basic principle of combustion analysis is the reaction of oxygen with carbon and sulfur compounds within the sample and quantifying the resultant CO₂ and SO₃ gasses. Early methods of detection of CO₂ involved weighing the CO₂ absorbed on a substrate or various titrimetric methods [17]. Detection of CO₂ by infrared spectroscopy of the combustion gas started in the 1970s and is the predominant method of detection in modern-day automated CS combustion instrumentation. Modern combustion analysis instrumentation is essentially automated “wet chemistry”. A simplified schematic of this type of instrumentation is shown in Fig. 1. Other similar equipment can be used for this purpose, as an example, infrared solid carbon analyzers which typically have quartz cells and sucrose calibrations.

The sample is loaded into a ceramic crucible along with various accelerators and fluxing agents needed to drive the combustion reaction to completion. Oxygen gas is used as the carrier/reactant gas. Prior to analysis at the IR detector, the combustion gas is scrubbed of water and SO₃ and any partially combusted CO gas is converted to CO₂ in a secondary oxidation furnace [15]. As sources of carbon contamination are ubiquitous in labs, precautions should be taken to avoid contaminating samples in the preparation process. It is good laboratory practice to fire the ceramic crucibles at a minimum of 1000 °C in air prior to use and then store in a laboratory desiccator.

Specimens are tested in powder form, generally powdered below 150 μm, and weighing around 0.2 g. Prior to analysis, the specimen should be dried to provide a reproducible mass basis to calculate the CO₂ mass fraction. For accuracy of measurements, calibration must be performed using known, standard, high-purity materials with a wide range of CO₂ amounts.

There are very few, if any, reference materials available with the CO₂ mass fraction to cover the range of CO₂ measurements expected in cement-based and other construction materials. National Institute of Standards and Technology (NIST) is engaged in an effort to produce research grade reference materials to support upcoming interlaboratory

Fig. 1 Simplified schematic of a combustion analyzer showing carbon combustion products (adapted from [18])



studies for an ASTM test method for using combustion analysis for cements and concretes [19]. A validation method for CO_2 measurement is described in ASTM C114 [15], however, it seems suited to lower mass fractions of CO_2 .

Over the course of several days, repeatability measurements in a single lab, single operator study yielded about 2% relative repeatability on standard reference materials [19]. The measured values of mass of carbon have to be converted to mass of CO_2 , based on stoichiometry; further corrections are needed based on existing carbon if uptake should be determined. We note that such conversions and corrections are also needed for TGA and other methods.

Combustion instrumentation is widely available in many laboratories thanks to their moderate cost, the ease of sample preparation and the speed of analysis. It is arguably, together with TGA, the most commonly used method to determine mineralized CO_2 content. Modern instrumentation can analyze a single sample in about 2 min or less, conferring the method with a distinct speed advantage over TGA.

As the main limitation, combustion analysis has no direct capability to determine carbon speciation. All of the carbon in a completely combusted sample will be converted into CO_2 gas regardless of its initial form. As such, if it is desired to measure the CO_2 uptake of a mineral carbonation product, it is necessary to make measurements before and after the carbonation steps. It should be noted that the stoppage of hydration and preparation of the sample by using isopropanol or other organic material may cause quantification error for combustion and other methods [10, 20]. Differentiation of mineralized and other forms of carbon, including organic carbon from admixtures, additives or incomplete pre-combustion e.g. in fly

ashes, is not possible in combustion-based methods as-is. However, it is possible to add additional sample preparation steps to differentiate the carbonate content from organics by acidifying the sample to drive off CO_2 mineralized in the form of carbonates [21]. Differentiation of CaCO_3 and MgCO_3 or other metal carbonates is not possible in combustion methods. Similar to TGA, the method is destructive.

4 Solid-state nuclear magnetic resonance spectroscopy

Solid-state nuclear magnetic resonance (NMR) spectroscopy, employing magic-angle spinning (MAS), has been widely used in studies of cement-based materials [22–25], where it has been utilized as the method probes one nuclear spin isotope at a time and because amorphous and crystalline phases are both detected in an equal manner. ^{27}Al and ^{29}Si NMR can be used to assess various changes in the cement paste over the carbonation process (Fig. 2), such as decalcification of the C-(A)-S-H phase, structural studies of the formed alumina-silica gel including aluminum coordination, and investigations that probe the pozzolanic reactivity of the alumina-silica gel in new cement formulations [26, 27].

^{13}C MAS NMR allows detection of the calcium carbonate polymorphs, calcite, aragonite, and vaterite, which give ^{13}C resonances at 168.8, 171.2, and 168.7 ppm, respectively [28, 29], as illustrated in Fig. 2 by the detection of calcite in a carbonated cement paste. Moreover, $^{13}\text{C}\{^1\text{H}\}$ cross-polarization (CP) NMR experiments, where magnetization is transferred from ^1H to ^{13}C , allow detection of hydrous carbonate species, e.g. on the surface of



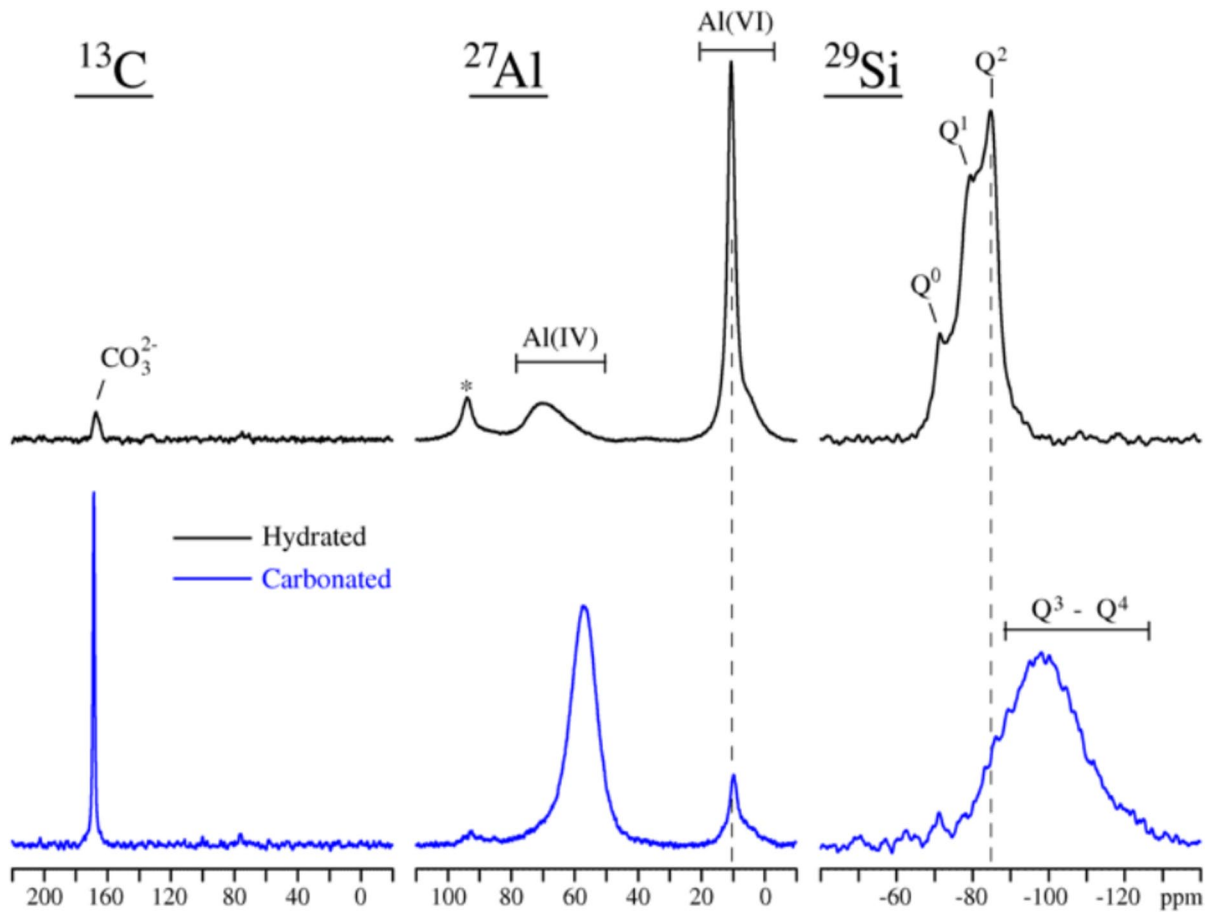


Fig. 2 ^{13}C , ^{27}Al , and ^{29}Si MAS NMR spectra of a well hydrated Portland cement paste before and after aqueous carbonation for 360 min (reproduced from [26], CC BY 4.0)

CaCO_3 particles, and of amorphous calcium carbonate ($\text{CaCO}_3 \cdot n\text{H}_2\text{O}$) [30] despite its resonance is overlapping with the peaks from the crystalline CaCO_3 polymorphs.

Solid-state NMR analyses are generally performed on dry, powdered samples and require sample amounts of 25–250 mg, depending on the nucleus under investigation and the special setup in terms of attainable MAS spinning frequencies (*i.e.* NMR probes and rotor diameters) [25]. The method is non-destructive but also associated with low sensitivity. Thus, the low natural abundance of ^{13}C (1.1%) and ^{29}Si (4.7%), combined with their generally long relaxation times (T_1) implies that ^{13}C and ^{29}Si NMR experiments typically require overnight acquisitions to achieve acceptable signal-to-noise ratios. Another limitation of the method is the sensitivity to high

levels of paramagnetic ions (*e.g.* Fe), which will result in an efficient relaxation of magnetization that may prevent detection of the signals. This may be a challenge for carbonated materials containing large quantities of fly ash or steel slag with high levels of iron. An absolute quantification of CO_2 uptake in the form of CaCO_3 can be accomplished by solid-state ^{13}C MAS NMR, however, the experiments would be rather time consuming, considering the generally long T_1 relaxation time for ^{13}C in carbonate ions along with the low natural abundance of the isotope, among other reasons.

Magnesium carbonates have been studied using ^{13}C MAS NMR and a detailed understanding of magnesium carbonate phases and M-S-H in complex systems can be obtained by combining Fourier-transform infrared spectroscopy, XRD, and NMR

[31]. Solid-state ^{13}C NMR has been shown to be able to distinguish various magnesium carbonate phases, such as magnesite, hydromagnesite, dypingite, and nesquehonite [32].

5 Fourier-transform infrared spectroscopy

Fourier-transform infrared spectroscopy (FTIR), which provides information about the different molecular species present in any sample, is a common infrared spectroscopy method. FTIR has been used to quantify depth of carbonation for concrete durability studies [33]. Combining XRD and FTIR, authors claimed to identify amorphous calcium carbonate in hydrated ternary cements and aragonite in the systems post-carbonation [34]. In studies evaluating the effects of carbonation and carbonation curing, authors identified using FTIR, microstructural modifications, including calcium hydroxide carbonation, C–S–H decalcification, and the formation of calcite, vaterite, aragonite, and amorphous calcium carbonates [35]. The progress of the enforced carbonation process, i.e. the increase in the calcium carbonate content, for various material streams could be followed in a qualitative manner using FTIR [36, 37]. FTIR can also be potentially used to quantify CaCO_3 amounts [38].

CaCO_3 , depending on the polymorph and phase, shows various IR absorption bands between 600 and 1500 cm^{-1} , as summarized in Table 2 [39–47] for the following vibrations: symmetric stretching (ν_1), out-of-plane bending (ν_2), asymmetric stretching (ν_3), and in-plane bending (ν_4). Figure 3 shows FTIR spectra for calcite, aragonite, and vaterite [40]. In simple (model) systems, the absence of ν_1 can be used to differentiate calcite from aragonite/vaterite, the shift in ν_4 to identify vaterite, etc., however, this is complex in cementitious systems due to overlap from

aluminosilicate and other bonds. Thus, in most studies, it is common to combine FTIR with complementary techniques, including XRD, TGA, and Raman, among others. FTIR absorption bands of CaCO_3 polymorphs can vary slightly depending on measurement conditions, instrument calibration, and sample characteristics, further highlighting the importance of using multiple techniques.

The most common FTIR types are transmission FTIR and ATR-FTIR (attenuated total reflectance mode). For the first method, sample preparation for solids requires KBr pellets. ATR-FTIR is an attractive measurement, which is rapid (measurement only takes 1–2 min) and does not require any sample preparation beyond a clamping arm that applies pressure on the powder. Powder is extracted from the paste, mortar, concrete, ground to cement or similar fineness, and tested using FTIR. Less than 0.5 g of powder is adequate for FTIR testing. Issues with sampling and aggregate interference are similar as with TGA and other powder methods, and discussed later. FTIR spectra of raw materials can help understand and minimize interferences from pre-existing CaCO_3 .

The main limitation of FTIR is that, it is for the most part, only qualitative. Through careful processing, calibration, and accounting for matrix effects, quantification in cementitious materials was demonstrated in a recent study [38], and further studies could advance FTIR quantification. Poor signal-to-noise, weak peaks, and peak overlap are other major limitations. These could be overcome through careful sampling, selective dissolution, subtractive comparisons, and combining data with other complementary techniques. The test can be destructive or non-destructive, depending on exactly how it is performed.

The (ν_2, ν_3, ν_4) modes of calcite, dolomite, and magnesite can be differentiated [48], and therefore,

Table 2 Band positions (cm^{-1}) and assignments of common CaCO_3 phases in cementitious systems [39–47]. Peaks which are typically weak are identified with a ‘w’

| Phase | Composition | ν_1 | ν_2 | ν_3 | ν_4 |
|-----------------------------|---|---------|---------|---------|-----------|
| Calcite | CaCO_3 | - | 874 | 1420 | 713 w |
| Aragonite | CaCO_3 | 1083 | 844/854 | 1461 | 700/712 |
| Vaterite | CaCO_3 | 1087 w | 876 | 1476 | 744 w |
| Amorphous calcium carbonate | $\text{CaCO}_3 \cdot x\text{H}_2\text{O}$ | 1067 | 864 | 1492 | 690/725 w |
| Monohydrocalcite | $\text{CaCO}_3 \cdot \text{H}_2\text{O}$ | 1063 w | 872 | 1492 | 698 |
| Ikaite | $\text{CaCO}_3 \cdot 6\text{H}_2\text{O}$ | 1085 w | 876 | 1425 w | 720 w |



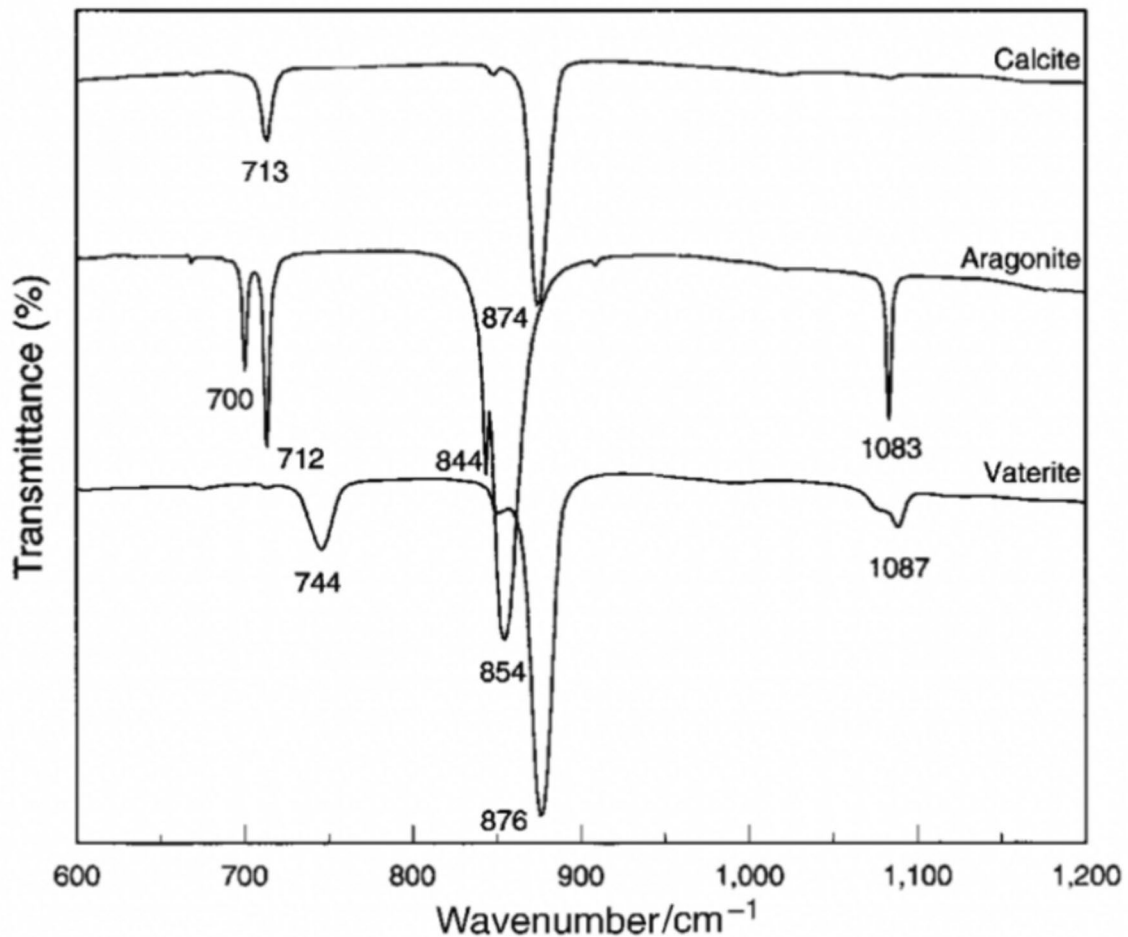


Fig. 3 FTIR spectra for calcite, aragonite, and vaterite (reproduced from [40])

these minerals can be differentiated in FTIR, at least in simple systems. FTIR has also been used to distinguish different members of the siderite-magnesite solid-solution series using shifts in the 1430 cm^{-1} and 880 cm^{-1} bands [49].

6 Raman spectroscopy & imaging

Raman spectroscopy and imaging have emerged as valuable tools for assessing the carbon uptake in cementitious systems. They have been widely used to measure the carbonation level at different depths [50], obtain the composition of carbonated cement paste [51], and investigate the biomineralization processes involving amorphous calcium carbonate

[52]. Various polymorphs of calcium carbonate formed during carbonation can be identified using Raman spectroscopy [53, 54]. Table 3 summarizes band positions for common carbonate phases. As with FTIR, some variability in the exact positions can occur.

Calcium carbonates have a strong Raman scattering between 1080 and 1090 cm^{-1} (ν_1 peak), corresponding to the symmetric stretching of C-O bonds [55]. Additional (ν_2, ν_3, ν_4) modes provide further insights into different calcium carbonate species including amorphous phases [52, 56]. Figure 4 illustrates the unique Raman spectra of some of the common calcium carbonate phases. Hence, Raman spectroscopy's high sensitivity to crystalline and amorphous phases enables accurate phase

Table 3 Band positions (cm^{-1}) and assignments of the carbonate phases identified in cementitious systems [52, 55–58]

| Phase | Composition | ν_1 | ν_2 | ν_3 | ν_4 | LV |
|-----------------------------|---|---------|---------|---------|---------|-------|
| Calcite | CaCO_3 | 1085 vs | – | 1450 vw | 712 w | 280 m |
| Aragonite | CaCO_3 | 1085 vs | 853 vw | – | 700 vw | 205 w |
| Vaterite | CaCO_3 | 1074 m | – | – | 740 vw | 300 w |
| Amorphous calcium carbonate | $\text{CaCO}_3 \cdot x\text{H}_2\text{O}$ | 1077 s | 868 vw | 1390 vw | 698 w | – |
| Monohydr-ocalcite | $\text{CaCO}_3 \cdot \text{H}_2\text{O}$ | 1069 m | – | – | – | – |
| Ikaite | $\text{CaCO}_3 \cdot 6\text{H}_2\text{O}$ | 1072 vs | 873 w | – | – | – |

vs: very strong, s: strong, m: medium, w: weak, vw: very weak

identification at high resolution [50]. Moreover, Raman imaging provides valuable spatiotemporal information on the carbonation of structural components in concrete (Fig. 4).

Capturing images at different locations and time points, Raman imaging allows an understanding of the distribution, depth, and extent of carbonation in concrete structures. While Raman imaging itself is generally non-destructive, the core of the concrete is required to assess the carbonation profile across the cross-section of the concrete [50].

Several studies have successfully employed Raman spectroscopy and imaging to monitor and quantify carbonation in cementitious systems [50, 52, 59, 60].

For instance, a portable Raman spectrometer has been used to study the kinetics of carbonation of portlandite, enabling the detection of the carbonation front through principal component analysis [60]. Raman spectroscopy has been employed for monitoring the evolution of several phases during natural carbonation [53]. Additionally, the correlation between carbonate content and Raman peak intensity has demonstrated a good agreement with conventional techniques [50]. Furthermore, Raman imaging has been employed to monitor the spatial and temporal evolution of calcite and portlandite in cement pastes over 14 days on a large area ($1 \text{ mm} \times 1 \text{ mm}$) [59].

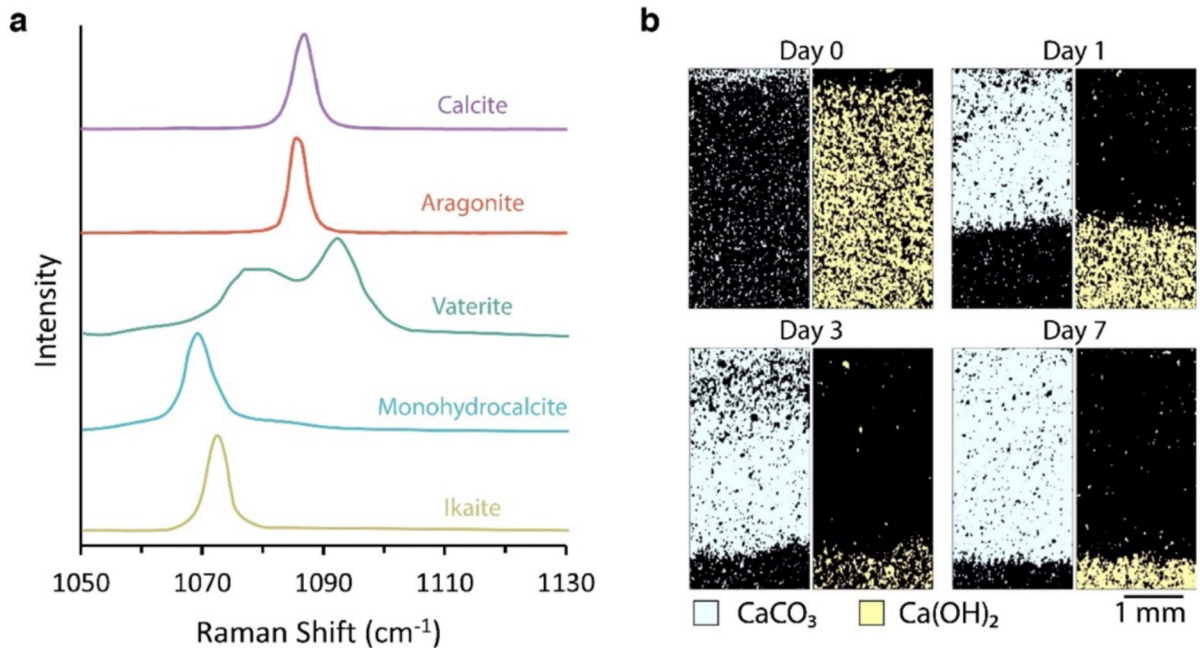


Fig. 4 **a** Raman spectra of different CaCO_3 phases, calcite, aragonite, vaterite, monohydrocalcite, and ikaite (adapted from [54]); **b** Raman phase maps ($1.8 \times 3.6 \text{ mm}$) of calcite and port-

landite on a hydrated cement paste specimen (w/c of 0.4) subjected to carbonation (adapted from [59])



To obtain high-quality data, it is preferred to use polished samples with low roughness [61]. Complications in data analysis can arise due to poor signal-to-noise ratio and overlapping peaks. However, these challenges can be mitigated by optimizing the scan parameters and performing thorough spectral analysis. Challenges that warrant further investigation include the long duration of scan time for accurate quantification via large-area scans [62], other scanning parameters, and the selection of an adequate representative area for analysis. Moreover, further investigation on the quantification methods for the carbon uptake capacity of various cementitious systems over a representative area is required.

Magnesite, similar to calcite, shows (ν_1, ν_3, ν_4) bands, and ν_4 of magnesite, at 738 cm^{-1} is significantly different from that of the 712 cm^{-1} band of calcite to enable differentiation [63]. Raman spectroscopy has also been shown to be able to differentiate various members of the siderite–magnesite anhydrous solid solution series [64]. Hydromagnesite and other phases have been studied using Raman spectroscopy [65], however, disorder, ionic substitution, and co-existence of phases make analysis and obtaining high quality spectra using solely Raman challenging. Combining Raman analysis with complementary techniques such as XRD is recommended for unambiguous identification and quantification. Recently, Rathnakumar and Garg [66] conducted multi-modal imaging on the carbonation front where multiple techniques like laser profilometry, contact angle goniometry, XRD, and TGA were combined with Raman Imaging to record evidence of pore refinement upon carbonation [66].

7 X-ray diffraction

X-ray diffraction (XRD) is an analytical technique widely used to identify and quantify crystalline phases in powders. In quantitative phase analysis, the presence of amorphous phases is taken into account by comparison to internal or external standards or by external calibration. Similar to other powder methods, the measurement of CO_2 uptake by XRD is indirect and based on the quantification of individual crystalline carbonate containing phases. An important limitation and bias of XRD analysis is that CO_2 in amorphous phases is not detected. Typically, XRD is

able to detect and quantify crystalline phases present in levels above 0.5–1 wt.% in multiphase mixtures. Multi-laboratory precision of XRD is at the 1 wt.%-level in favorable cases, but may increase in complex mixtures such as Portland cement to up to 5 wt.% for major phases. No dedicated studies on determination of precision or bias of CO_2 uptake measurements by quantitative XRD have been reported.

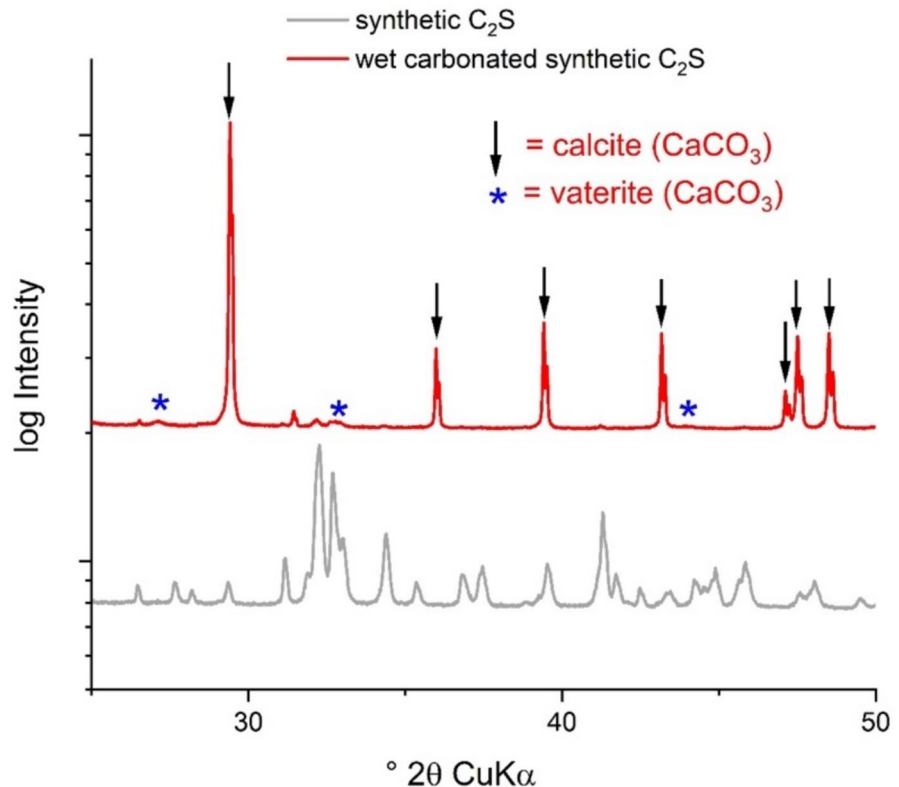
The measurement principle of XRD is based on the elastic interaction of X-rays with the electron density of atoms situated in a crystal lattice. The diffraction pattern is mainly determined by the symmetry and dimensions of the crystal lattice, the type and the positions of atoms in the unit cell. Using the characteristic combination of peak positions and intensities, crystalline phases can be identified. The well-known Rietveld method uses and refines crystal structure information to fit a calculated pattern to a measured powder diffraction scan. In addition, it is widely used to measure phase abundancies in multi-phase materials, such as cements or carbonated materials.

Figure 5 shows an example of aqueous carbonated C_2S where XRD enabled to detect calcite and vaterite as the main carbonate phases. This clearly highlights the complementarity of XRD to other techniques that cannot distinguish among different CaCO_3 polymorphs.

The identification and quantification of carbonate phases formed strongly depends on the raw material. In single phase systems, the identification and quantification of carbonate phases is straightforward without problems caused by peak overlapping. However, for many real materials, such as carbonated cement or steel slag, peak overlap may be severe and require analysis using full-pattern fitting. Figure S1 shows the XRD patterns of carbonated C_2S and carbonated steel slag. The identification of calcite and vaterite in the C_2S system is quite clear, while the identification of the calcite and Mg-calcite peaks in the steel slag system is much more challenging due to the multiphase nature of the raw material containing phases including bredigite, cuspidine, donathite, and more.

Powder XRD is a destructive method. It is mandatory that the particle size distribution of samples is very fine, preferably below $10\text{ }\mu\text{m}$, to avoid bias and overestimation of amorphous content [68]. Consequently, the sample has to be milled while taking care that the material is not altered by the milling action. For a typical measurement in reflection mode, around

Fig. 5 Synthetic C_2S aqueous-carbonated in 100% CO_2 atmosphere at atmospheric pressure and with excess of water (using data from [67])



2 g of sample are needed. As with other powder methods, care should be taken that the sample measured is representative of the material to be tested.

Identification of crystalline phases or qualitative interpretation of powder XRD data is straightforward using commercial or free software in combination with suitable identification databases. Built-in search/match routines can be combined with a custom-made list of carbonate phases that are likely present. In principle, XRD can be used to detect carbonation reactions in a range of binder systems, since both crystalline reactants (e.g., portlandite, C_2S , tobermorite) and crystalline products (e.g., calcite, aragonite, vaterite) can be identified and measured. Qualitative XRD analysis showed that an investigation of carbonation depth using phenolphthalein in concrete yielded depth values about half as large as carbonation reaction products detectable by XRD [33]; however, other studies showed similar depths [69, 70]. Qualitative XRD was also used to study qualitatively the enforced carbonation of cement fines by tracing the formation of crystalline calcite during carbonation [59, 71, 72].

Quantitative analysis of XRD data is more challenging and commonly done using the Rietveld method implemented in a range of software packages. As a result, individual phase abundancies are obtained. As amorphous phases can be present initially or formed during the carbonation processes, quantitative XRD analysis should not ignore these for the phase contents to be reliable. To quantify the amounts of amorphous phases, either internal standards [37, 68] or external standards approaches [73, 74] are commonly used. If the external standard method is chosen, the chemical composition of the samples, including the H_2O and CO_2 contents, are needed [74]. This may preempt the need for a measurement of CO_2 uptake by XRD. However, since the difference in X-ray absorption between CO_2 and H_2O is not significant, a simple measurement of loss on ignition can be used as well to calculate the sample mass attenuation. This information and calculation are not needed in the internal standard approach, that, in turn, requires careful mixing and homogenization of the chosen internal standard into the material. When operating in transmission geometry, where a

well-defined sample volume is irradiated, an internal standard should be used, because achieving exactly the same packing density for the external standard and the sample is practically not feasible.

In Bragg–Brentano diffractometry, however, an external standard can be applied. In this geometry, the effective irradiated sample volume varies with the diffraction angle, and—due to the comparatively high mass attenuation coefficients of cementitious materials—the specimen is not fully penetrated down to the sample holder. Since air is essentially “invisible” to X-rays compared to the solid phases, differences in packing density between the external standard and the specimen do not play a decisive role in this case.

Several studies have been published using quantitative XRD to quantify enforced carbonation of materials. In [75], hydrated cement paste was exposed to enforced carbonation for up to 30 h and the evolution of different mineral phases was determined by Rietveld analysis (Figure S2). For determination of the amorphous content, the external standard method (rutile) was used. Salman et al. [76] used the internal standard method (with ZnO) to determine the phase composition including the amorphous phase content of a steel slag carbonated under different conditions. Other studies have also used corundum (Al_2O_3) as the internal standard method for amorphous phase determination [77, 78]. As stated in [79], ZnO can be preferred over corundum due to stronger reflections and the absence of X-ray amorphous material.

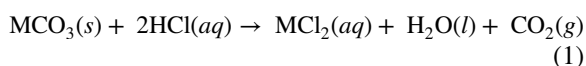
If calcite is formed during carbonation of Mg/Fe containing systems, a Mg/Fe containing calcite can be formed. Solid solutions of Mg/Fe can be identified by lattice parameter shifts of the calcite formed [80]. During Rietveld refinement several other issues must be considered such as preferred orientation, over-grinding and peak broadening during milling of the samples [81]. Table S1 presents a list of carbonate phases commonly formed during carbonation processing as well as recommendations concerning the ICSD structures to be used for Rietveld refinement.

8 Chemical methods

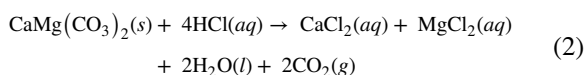
Chemical methods for quantifying CO_2 are based on reactions between the material and a reagent, allowing for estimates of the CO_2 content. In case of aqueous

carbonation, it may be necessary to determine the initial CO_2 value by calculation [82].

The most commonly used chemical method is based on reacting the carbonates with acid and assessing the amount of CO_2 released; this is known as calcimetry (or gasometric analysis). For a mineral containing metal carbonates, the reaction with HCl is governed by the following rapid reaction Eq. 1 (where M equals Ca, Mg, or Fe).



The acid digestion reaction is applicable for Portland cement, blended cements, magnesium-based cements, and other materials. In some cases, a secondary and slower reaction occurs. An example for the acid digestion for dolomite is shown in Eq. 2:



In calcimetry, the volume of CO_2 released can be quantified, when the test is carried out at constant atmospheric pressure using the Bernard or Dietrich-Frühling calcimeters. Alternatively, the pressure variation can be measured when the test is carried out at constant volume using a carbonate bomb setup (or carbometer). Such tests are commonly used by geologists to determine the purity of limestone [83, 84]. Tests need to be performed on milled samples of around 1 to 2 g, and are destructive. The calcimeter test, illustrated in Fig. 6a, is specified by the French standard NF P 94–048 [85] and the British Manual of Soil Laboratory Testing [86]. The “carbonate bomb” apparatus (Fig. 6b) is based on the work of Müller [87]. The carbonate bomb requires an initial measurement of a CaCO_3 reference standard.

Chemical methods can be based as well on a gravimetric principle where the amount of CO_2 is measured by total weight loss after reaction with acid, assuming the conservation of produced water [89]. In that case, tests can be performed on milled or crushed samples of around 10–30 g. It is a destructive method as the sample must be extracted from the carbonated material [90, 91]. The test is performed by stirring the solution in a conical flask covered with a watch glass and weighing at regular intervals. To allow the volatilized CO_2 to escape from the flask, the glass bell is removed for a few seconds before each weighing. The



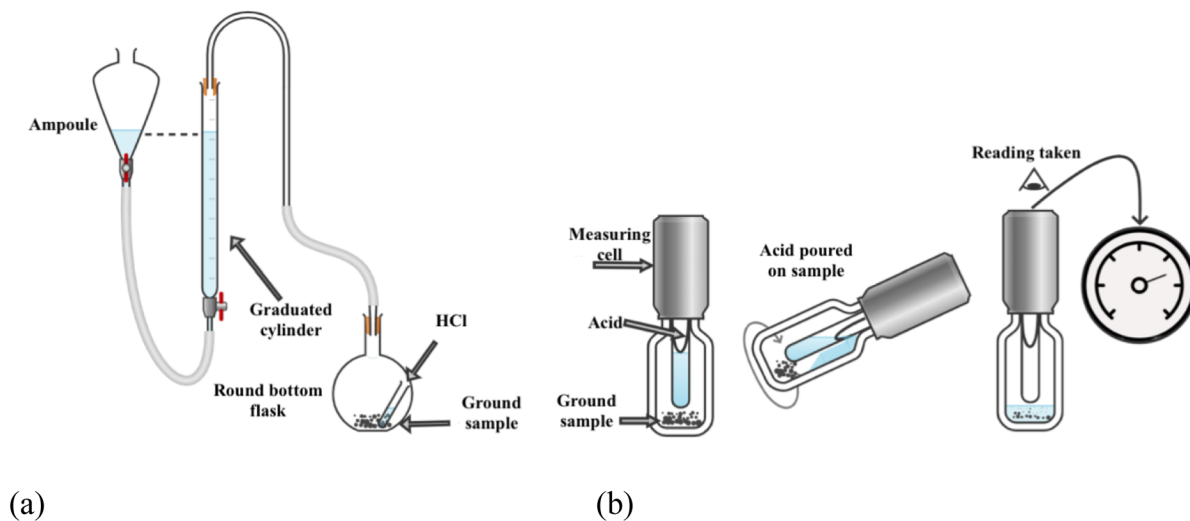


Fig. 6 **a** Bernard calcimeter, adapted from ([85, 86, 88]; **b** carbonate bomb (adapted from [87])

test is carried out until the weight loss between two consecutive values is negligible [92]. As silica gel may hinder the dissolution of encapsulated carbonates, finer samples may have to be used [10]. The test has been used recently to assess CO₂ sequestration in aqueous carbonation of electric arc furnace slag and cement by-pass dust [93, 94].

The digestion-titration method uses a two-step acid-digestion and titration method with a simplified digestion apparatus wherein, at reduced pressure, an acid is added to liberate the chemically bound CO₂ in the specimen (Eq. 1) which is then consequently dissolved in a pre-prepared Ba(OH)₂ solution. Following digestion, titration and back-estimation of the specimen CO₂ content are performed using an end-point indicator (e.g., phenolphthalein) and the chemical reaction stoichiometry of the overall process. This method was first developed in the early 1900s by Van Slyke for clinical biochemical studies, but is applicable to carbonates in general [95]. Thien et al. [96] made a slight modification to the specimen holder to avoid errors associated with excessive bubbling during digestion. The apparatus is shown in Figure S3.

Acid digestion can also be used in association with other methods for CO₂ determination. For example, the acid reaction can be coupled to a non-dispersive infrared analyzer to quantify total inorganic carbon (TIC). Tests are performed on 50–100 mg of carbonated samples (<160 μm) digested with

orthophosphoric acid [97] or other acids and then heated to 200 °C for determination of TIC in the evolved gas flow.

The acid concentration in the tests is chosen so as to guarantee an excess of acid during reaction with MCO₃ but also with the other components of the mineral. Temperature and pressure corrections, or calibration with a calcite standard, are required for calcimeters [98]. The accuracy of the volume reading of the calcimeter is ±0.25 ml CO₂ released, equivalent to 1 mg CaCO₃. The accuracy of the pressure reading of the calcimeter expressed as carbonate content is ± K*0.5% CaCO₃, where K is the calibration coefficient [99]. For gravimetric estimates, the weight of the released CO₂ is estimated with an accuracy of ± 0.01 g. For the digestion-titration method, the accuracy of the method is directly related to the titrant drop size during titration. Baseline titrations help improve the accuracy of results [96].

As part of the FastCarb project [98–101], recycled concrete aggregates were carbonated on an industrial scale and the CO₂ uptake was determined by calcimetric methods which were found to be accurate for estimating the process efficiency of the carbonation process. They offered average repeatability, a good reproducibility for the calcimeter, and even better for the carbonate bomb. The calcimeter was more accurate when samples were tested in the laboratory as opposed to the carbonate bomb which is more suitable for on-site testing [98, 99].



In many cases, it is assumed that the temperature is kept almost constant while testing. If this is not the case, cooling is required to compensate for the highly exothermic reaction [98, 99]. In the case of high CO₂ levels, sampling is reduced to ensure an excess of acid,

CO₂ is not stored in these phases. While hydrated magnesium carbonate can form by carbonating MgO, this phase is not observed after carbonating PC paste; rather, MgO is stabilized in hydrotalcite [104]. Taking these factors into account, Eq. 3 can be rewritten as Eq. 4.

$$\text{TheoreticalCO}_2\text{uptake(wt\%)} = \frac{M_{\text{CO}_2}}{M_{\text{CaO}}} \left(\text{CaO} - \frac{M_{\text{CaO}}}{M_{\text{CaCO}_3}} \text{CaCO}_3 - \frac{M_{\text{CaO}}}{M_{\text{SO}_3}} \text{SO}_3 \right) \quad (4)$$

which may result in an insufficiently representative unit volume of the sample if its composition is heterogeneous. The presence of sulfides can produce a parasitic reaction, and filtering and other solutions are required to overcome this issue. A major concern for the carbonate bomb is that it is not adapted for samples with a carbonate content of less than 25% because the calibration coefficient of the carbonate bomb is less reliable for this measurement range. Detailed studies of other calcium carbonates, magnesium, and iron carbonates using calcimetric methods appear unavailable.

9 Carbon uptake and calculations

Reliably quantifying the climate change mitigation potential of mineral carbonation technologies requires accurate measurement of the CO₂ uptake of mineral carbonation products and the correct use of the results in rigorous life cycle assessment (LCA) studies.

The theoretical CO₂ uptake (or binding capacity) of a material is determined by the availability of alkali earth and alkali elements for carbonation. It is commonly calculated using the Steinoor equation based on oxide mass % and molar mass, shown in Eq. 3 [102].

$$\text{TheoreticalCO}_2\text{uptake(wt\%)} = \frac{M_{\text{CO}_2}}{M_{\text{CaO}}} \left(\text{CaO} - \frac{M_{\text{CaO}}}{M_{\text{CaCO}_3}} \text{CaCO}_3 \right) + \frac{M_{\text{CO}_2}}{M_{\text{MgO}}} \left(\text{MgO} - \frac{M_{\text{MgO}}}{M_{\text{MgCO}_3}} \text{MgCO}_3 \right) + \frac{M_{\text{CO}_2}}{M_{\text{Na}_2\text{O}}} \text{Na}_2\text{O} + \frac{M_{\text{CO}_2}}{M_{\text{K}_2\text{O}}} \text{K}_2\text{O} \quad (3)$$

Equation 3 accounts for the wt.% of CaO, MgO, Na₂O, and K₂O available for carbonation and includes the oxide/carbonate molar masses to compute the theoretical (maximum) CO₂ uptake. However, the actual amount of CO₂ uptake varies depending on the waste and carbonation sources used. Much data [26, 103, 104] shows that calcium sulfate will not carbonate at ambient temperature. Na₂CO₃ and K₂CO₃ are soluble, meaning

When pressurized CO₂ is used, carbonation of CaSO₄ [105] and MgO can occur, even though MgO carbonation is usually limited [106]. In this case, the equations above can be modified through the reconsideration of the SO₃ and MgO terms.

When powders are being carbonated, the degree of carbonation can be defined as the difference between the sample weights before and after carbonation with respect to the initial mass of the adopted dry powder. This method is suitable for carbonation of compact samples in CO₂ chambers. However, water release associated with the carbonation reaction should be taken into account and corrected for; a conservative estimate using masses before and after carbonation and the lost mass is shown in Eq. 5 [107, 108].

$$\text{CO}_2\text{uptake(\%)} = \frac{\text{Mass}_{\text{aftercarb.}} + \text{H}_2\text{O}_{\text{lost}} - \text{Mass}_{\text{beforecarb.}}}{\text{Mass}_{\text{beforecarb.}}} \quad (5)$$

In many cases, the CO₂ assessment is directly executed on a sample extracted from the mineralized product rather than during the CO₂ absorption process, so the initial mass of the sample before the carbonation is unknown and some assumptions are required to define the CO₂ uptake. In this case, the

CO₂ uptake can be related to the total mass of the carbonated sample [97]. For assessing CO₂ uptake using thermal treatment such as TGA, the initial mass before carbonation can be assumed as the mass of the carbonated sample after treatment at specific temperature values, such as 105 °C [109, 110], 800 °C [111], or 950 °C [112]. The use of temperature values higher than the decomposition temperature of



calcium carbonate (~ 750 °C) is more appropriate as it provides an estimate of the initial mass after the release of the CO_2 content. In many studies the initial mass before the carbonation is defined as the difference between the carbonated mass and its CO_2 content, as shown in Eq. 6 [97, 106], considering the CO_2 contents of the carbonated (final) and raw (initial) materials. This approach exclusively considers weight changes induced by CO_2 uptake. It is important to note that the measurement of CO_2 uptake can be subject to under- or overestimation if concomitant reactions during carbonation, including dehydroxylation or hydration, induce significant mass changes [109].

$$CO_2\text{uptake}(wt\%) = \frac{CO_{2,final}(wt\%) - CO_{2,initial}(wt\%)}{100 - CO_{2,final}(wt\%)} \quad (6)$$

A key parameter to quantify the carbonation is the carbonation efficiency (or the carbonation degree), which is defined as the ratio between the captured mass of CO_2 and the theoretical CO_2 sequestration potential assessed using the Steinoor equation [113].

Measurement of CO_2 uptake can be executed at the industrial scale, although few methods have been reported for estimating the CO_2 uptake of ready-mix concrete and precast concrete products cured with CO_2 . This can be achieved by measuring the amount of CO_2 injected into the curing system during the production of concrete. The amount of CO_2 lost during curing needs to be measured in this method. The amount of sequestered CO_2 in concrete can be calculated by subtracting the total CO_2 lost from the total amount of CO_2 injected into the curing system. However, measuring the amount of CO_2 lost can be challenging as leakage can occur at many locations. Therefore, the assumption that the amount of CO_2 that is sequestered in the concrete is equal to the total amount injected minus the total amount lost can be significantly overstated and ideally sequestered amounts should be verified/validated by measurement in the hardened concrete product after curing. Direct measurement of uptake can also be done with CO_2 concentration measuring probes placed in the carbonation chamber. The uptake rate is evaluated by the concentration drop and the total amount of CO_2 uptake by integration of this rate. Such methods have already been proposed for other applications such as monitoring the atmospheric carbonation of recycled

concrete aggregates [12, 114, 115]. A direct measurement of the uptake has the advantage of being non-destructive and allowing monitoring over time.

10 Using CO_2 uptake values in life-cycle analysis (LCA)

Two key issues in LCA studies of CO_2 utilization are additionality and temporality [116]. Since CO_2 is naturally absorbed and emitted between different natural reservoirs, a CO_2 utilization process should only be credited (i.e. to reduce CO_2 emissions) when the permanently captured CO_2 (in thermodynamically stable carbonates) is additional to the natural carbon cycle [117]. This condition applies to CO_2 utilization in cementitious materials since here CO_2 becomes chemically bound and is thermodynamically stable at ambient conditions. This additionality condition is met irrespective of whether the source of the CO_2 absorbed originates from fossil or geogenic sources, which would have otherwise increased the atmospheric CO_2 concentration; or from the air (through direct air capture or from biogenic carbon sources), which lowers the CO_2 already existing in the atmosphere [118]. While details of regulations are out of the scope of the article, EU regulations permit considering carbonation as a permanent means of carbon storage. As an example, regulation 2018/2066 [119] on the EU Emissions Trading System (ETS) permits the subtraction of CO_2 from reported (fossil) emissions when it is used to produce precipitated calcium carbonate, where the CO_2 is chemically bound. Similarly, EU Regulation 2024/3012 [120] on the voluntary certification of carbon removals, which covers atmospheric or biogenic CO_2 sources, recognizes “permanently chemically bound carbon in products” in its “permanent carbon removal” definition.

Crediting CO_2 absorbed through natural air carbonation of a cementitious material requires special consideration. This includes the significant variation in CO_2 uptake depending on its end use (e.g. indoor building decking, outdoor paving) and due to the uncertainty imparted by the temporality of natural carbonation that can span many years or even centuries (non-technical reasons significantly influence the in-use lifetime of concrete especially in buildings). Guidelines for natural recarbonation accounting are



given in EN 16757 [121]. Due to the urgency of climate change mitigation and the nature of the global warming potential of CO₂ gas which decays over time, it is more beneficial to reduce CO₂ emissions in the present than in the future [122]. Dynamic LCA, which provides a cumulative snapshot of the climate change impact over the product life cycle, is useful for studying natural carbonation especially when the time horizon for carbonation is long (as it is in natural recarbonation), but is more complex than static LCA [123, 124], and is not necessary to perform an insightful LCA study, especially for products of enforced carbonation.

CO₂ chemically bound by the cementitious material should always be included in the system boundary. The system boundary can be expanded to include the CO₂ source, or it can be modelled via a substitution allocation approach that substitutes flows of equivalent functionality, e.g., a product made from waste CO₂ mineralization substituting one made from virgin resources [125, 126]. Substitution-based allocation can hence credit the use of wastes as avoided impact from their alternate uses, e.g., landfilling. The definition of the system boundary depends on the goal and scope of each study. RILEM TC 309-MCP focuses on several LCA aspects including on carbonated recycled concrete aggregates [127].

The functional unit selected, typically a unit (e.g. 1 kg, 1 m³) of binder, aggregates, or concrete, should reflect the function of the cementitious material produced from the CO₂ utilization process [128]. For example, Ravikumar et al. [129] used a functional unit that includes volume and 28 days compressive strength in assessing ready-mix concrete and CO₂ injected ready-mix concrete. However, considerations of mass, volume, or strength ignore durability, or, at best, indirectly account for it. Bulk resistivity could possibly be used as an indirect but rapidly measurable indicator of durability [130–132]. However, the functional unit should also be simple and in general there is no perfect choice.

Primary data should ideally be used for the specific unit processes within the main scope of the LCA study, including upstream processes such as the capture, separation, transportation, and enforcement/injection of CO₂ [26]. Indirect sources of impact such as the required infrastructure, the maintenance and monitoring of the CO₂ utilization processes, and potential CO₂ leakage during transportation, can be

considered but may not be available or significantly affect the LCA study conclusions [129]. Quantities of CO₂ absorbed in the cementitious materials output by CO₂ utilization processes can be measured experimentally or modelled, as discussed earlier in this Section.

Although consideration of multiple impact categories provides a more comprehensive LCA result, it is usually justified in CO₂ utilization applications to focus on climate change, and report results at the midpoint level (global warming potential). Large uncertainty is imparted to the LCA results through the choice of LCA study parameters, therefore, carefully considered and clearly documented modelling choices regarding how allocation and CO₂ uptake is modelled, and what the basis of comparison is in a comparative LCA study, is essential for reliable quantification of the climate change mitigation potential of mineral carbonation.

11 Practical considerations and comparisons of test methods

To assess the CO₂ uptake of a process, whether natural or enforced carbonation, representative samples should be analyzed before and after the process step, or during the process. Determining the CO₂ uptake is relatively simpler for powder SCMs, but as the sample heterogeneity increases as we scale to concrete, the complexity increases profoundly. For all powder test methods, two major challenges are associated with sampling and sieving. When it is desired to measure the CO₂ uptake in a solid sample, the carbonation extent may vary as a function of depth. In such a case, the CaCO₃ measured at the surface and the bulk may be very different. Thus, testing of multiple powders from one specimen may be required. Even at one point, multiple test repeats are likely needed for confidence in measurements. As one goes from paste to mortar to concrete, the material heterogeneity increases, while the amount of mineralized CaCO₃ reduces (dilution). Interference from the presence of CaCO₃ in the aggregates must be overcome, as the measurement of interest is the mineralized CaCO₃ in the matrix but not the CaCO₃ that is present a priori in the aggregates. After powdering, sieving is needed. Sieving may be done in two ways, one in which the aggregates are removed, and the other

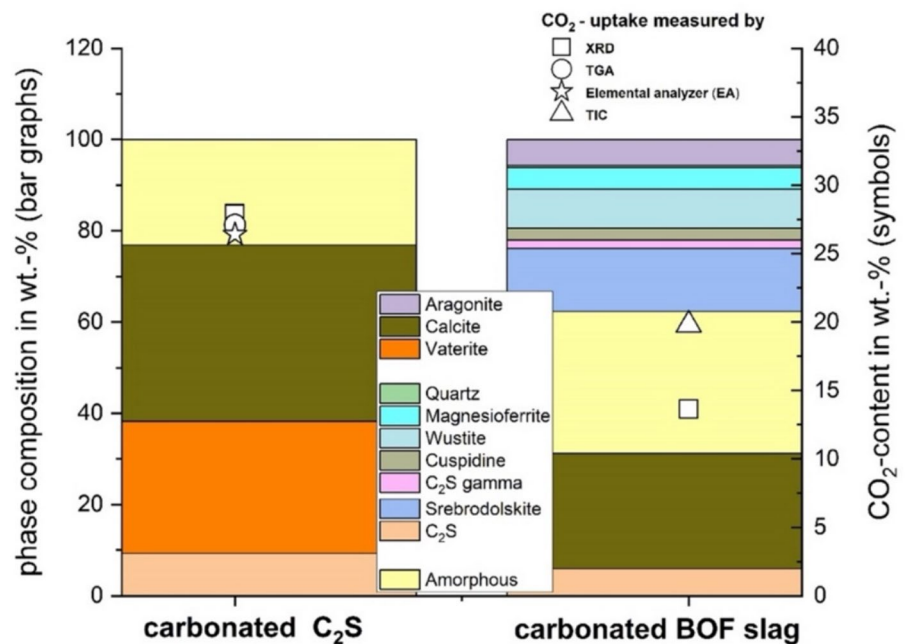
in which the entire material is ground down. In the first, incomplete removal of aggregates will cause the samples to have variable paste fractions. In the second, due to dilution, the amounts of mineralized CaCO_3 measured are low, and differentiation between the mineralized CaCO_3 and any CaCO_3 present in aggregates is needed. Either way, the major material sampling challenge is obtaining a representative powder in mortar/concrete without interference from the aggregates. This challenge becomes much more complex if the aggregates are limestone-based. When testing SCMs, cements, or carbonatable binders, aggregates are not a concern, however, there is interference due to the carbonates present in some binders. Thus, for accurate measurements, in addition to before/after measurements, the CO_2 content of all raw materials must be determined. When mix designs are unknown and raw materials are unavailable, determination of uptake becomes incredibly challenging, as appropriate corrections cannot be made.

In test methods where CaCO_3 amounts are determined, the CO_2 uptake is back-calculated. Generally, when tests such as XRD, TGA, LOI (from step-wise heating in a furnace), and combustion are compared, the amount/uptake results match well; however, careful consideration must be given to the limitations discussed previously. As an example, when using

XRD, there may be an underestimation of the amount of CO_2 for materials containing amorphous CaCO_3 . As seen in Fig. 7, the quantification of CO_2 content of a carbonated C_2S system corresponds well to the results of other independent methods such as TGA and combustion (denoted in the figure as elemental analyzer, EA). In contrast, the quantification of CO_2 content by XRD is underestimated after carbonation of a BOF slag as is apparent by comparison to results from other methods. This underestimation indicates the presence of amorphous calcium carbonates. Both examples show that XRD is instrumental in identifying different polymorphs of CaCO_3 ; FTIR and Raman can also aid with this differentiation. The BOF slag example also shows that combining XRD with TGA or other chemical techniques, enables to reveal the presence of amorphous calcium carbonate if care is taken to accurately determine the amorphous or unknown phase content. There is also the question of how the uptake should be expressed, whether normalization is by cement portion ($\text{kg CO}_2/\text{kg cement}$), overall concrete mass ($\text{kg CO}_2/\text{kg concrete}$), or by concrete volume ($\text{kg CO}_2/\text{m}^3 \text{ concrete}$).

Tran et al. [96] applied a modified Van Slyke's method to estimate the mineralized CO_2 content of construction materials utilizing two specimen sizes (i.e., 50 mg and 200 mg). To validate the method

Fig. 7 Phase composition and CO_2 uptake in a carbonated C_2S and BOF samples (replotted using data from [97]) showing the presence of different polymorphs of CaCO_3 , comparison of various methods, as well as the potential underestimation of CO_2 uptake with XRD



in a robust manner, digestion-titration results were compared to TGA results. Overall, the findings demonstrated that digestion-titration results were on par with TGA results for single-phase specimens, but slightly outperformed TGA for specimens composed of multiple phases.

Apart from sampling and sieving issues, for RCA samples, drying and crushing is also required; in such cases, the time for such preparation should be minimized. In all cases, if the carbonate content is not measured immediately, the sample should be stored in a dry place, away from all moisture and CO₂. TGA or combustion analyses are generally the ‘reference’ methods for quantification of carbonates or CO₂ content in cementitious materials. Obtaining accurate results and ensuring the representativeness of the sample requires repeat measurements to be taken since the sample size is small. For most samples, tests are done on the powders, with certain exceptions. Measurement on larger samples is challenging, but LOI up to 950 °C, can be used to quantify CO₂ content for large samples. Considering heat transfer and other parameters, the heating rate and holding time must be tailored to the specimen size. The error of this quantification method is typically 0.5% when done in well controlled conditions, determined for recycled concrete aggregate [99].

The Fastcarb project provided recommendations and compared methods for measuring captured CO₂

at industrial scales [98–101]. It was concluded that while TGA has its uses, LOI and calcimetry methods offer more reliable and practical options for measuring CO₂ in carbonation processes. TGA is suitable when changing material sources but is complex. LOI is ideal for characterizing average batches of materials before and after carbonation. It can be used for long-duration testing (12 h) on average recycled concrete aggregate batches. Calcimetry is effective for quality control between LOI measurements for production control to correct issues such as drift in carbonator settings [98, 99]. Studies have also shown limited differences between calcimeter, thermogravimetry, and gravimetric analysis when tests have been carefully performed [133]. In general, careful sample preparation and following appropriate procedures/guidelines is critical for obtaining quantitative data in cementitious material systems [134]. As this field grows and more advances are made, careful testing, quantification, and understanding of limitations of test methods [135] will remain critical.

Table 4 shows a comparison of the various test methods discussed in the text. Mass change and measurements of CO₂ injected are not listed; we note that these are likely the methods which are most applicable for industrial process control.

Table 4 Comparison of various carbonate/CO₂ uptake test methods. Quant is quantitative, dest is destructive. The * indicates generally but not always. Cost, in USD, is an estimated

cost of the equipment for a ‘base’ model. Given the vast variability in the parameters, the numbers listed to a certain extent are subjective

| Method | Measurement | Quant? | Dest? | Speed | Cost | Range/Error |
|------------------|--|--------|--------|----------|------------------|--|
| TGA | CO ₂ mass loss, can be converted to carbonates | Yes | Yes | 2 h | \$50,000 | <1% CaCO ₃ |
| LOI | Mass loss, can be converted to CO ₂ (or carbonates) | Yes | Yes | 2 h | \$5,000 | 0.5% Error |
| Combustion/CS | Carbon mass loss, can be converted to CO ₂ | Yes | Yes | 5 min | \$50,000 | <1% CO ₂ Repeatability <2% single operator |
| NMR | Information about polymorphs | No | No | 2 h | \$500,000 | – |
| FTIR | Peaks give info about polymorphs and amorphous CaCO ₃ | No* | No* | 5 min | \$50,000 | – |
| Raman | | No* | Yes/No | Variable | \$300,000 | – |
| XRD | Peaks give info about polymorphs and amounts | Yes | Yes* | 30 min | \$100,000 | Precision 1 – 5% |
| Chemical methods | Volume, pressure, or mass changes converted to CO ₂ | Yes | Yes | Variable | \$5,000 or below | 1 mg CaCO ₃ Variable |



12 Summary

All concrete uptakes CO_2 , as do other cement-based materials. Also, intentionally carbonated construction products use and permanently store CO_2 . Reliably quantifying the climate change mitigation potential of mineral carbonation products requires accurate measurement of their CO_2 uptake. TGA, LOI, combustion analysis, NMR, FTIR, Raman, XRD, chemical methods, mass change, measurement of CO_2 injected, can all be used to measure CO_2 (or carbonate) content. Of these methods, NMR, FTIR, and Raman are generally not quantitative. TGA and combustion analysis are the most commonly used for quantification purposes. At an industrial scale, combustion, mass change, and measurement of CO_2 injected are likely the only ones that are feasible. When using powder, representative sampling is critical; contamination/dilution from aggregates in concrete products or aggregates complicates measurements. To accurately determine CO_2 uptake, measurements before and after CO_2 exposure are needed, or, alternatively, measurements on raw materials. Through the use of complementary methods, as an example, the use of XRD to determine the CaCO_3 polymorphs, accurate quantification and an understanding of mechanisms can be obtained. Further round-robin testing, comparing the various methods is an urgent need to advance measurements, including in the lab, industrial scales, and in the field.

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Declarations

Conflict of interest Ellina Bernard, Ipei Maruyama, and Prannoy Suraneni are Associate Editors of Materials and Structures. The authors declare no competing interests in this publication.

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